INTRODUCTION

Currently, researchers focus on pollution control, especially wastewater treatment, to answer environmental control of textile, paper, petroleum, and agriculture manufacturing. The wastewater pollutants are mostly organic dyes, which contain high concentrations of organic pollutants. Aquatic systems may be exposed to wastewater with intense color and toxicity. The structure of the dye contains an aromatic ring that is harmful to the environment (Arumugham et al., 2018). Among them are the azo dyes with the azo group \(-\text{N} = \text{N}\), which are also recognized to be harmful for the environment due to toxic materials and their ability to cause cancer in humans and pollution in oceans, lakes, and rivers (de Almeida et al., 2021). The common techniques for treating dye wastewater include biological and chemical procedures (Sirianuntapiboon et al., 2007). Biological wastewater treatment when compared to other treatment options, is frequently the least expensive. However, it is thought that dyes are resistant to biodegradation (J. Zhang et al., 2020). Because the decay rates heavily depend on the reactivity and photosensitivity of a dye, direct...
photolysis of organic dyes in the natural environment has proven to be challenging (Neppolian et al., 2002). No treatment method now in use is 100 percent successful, and some call for the use of several different strategies.

The conventional treatment applied for wastewater pollutants, such as the adsorption process, coagulation, and membrane separation remains to be solved by the degradation problem. Effective alternative treatments are needed. Recent efforts included developing the degradation treatment based on the presence of UV irradiation. The last ten years have seen considerable advancements in the photocatalysis process, which has generated a great deal of attention. The photocatalytic process has generated a great deal of attention. The photocatalytic process also offers an intriguing benefit for this kind of pollution; in fact, the direct mineralization of the azo group is the perfect example of environmental treatment, because it results in the synthesis of an element that makes up the atmosphere.

The dye in wastewater can be eliminated by UV irradiation and nano-sized TiO$_2$ powder (Ali et al., 2021). TOC is the most effective method of cleaning up wastewater contaminated with procion red mx-5b (Cotillas et al., 2018) (Figure 1). EDTA-modified BiFeO$_3$ reached 70% degradation of procion red (Da Cruz Severo et al., 2020). Lin and Lee. (2010) TiO$_2$/Ag materials that have been created demonstrate effective photocatalysis. At the same time, Costa et al., (2004) achieved 93% resulting color for the red and yellow dyestuffs by the photodegradation process.

Among all the various materials, LDH has been promoted for photocatalyzed pollutant degradations because of its stability, large surface area, adjustable band gap, remarkable recyclability, and high anion exchange capacity (Zhang et al., 2019). Layered double hydroxides (LDHs) are ionic layered structures with the general formula $[\text{M}^{\text{II}}{}_{1-x}\text{M}^{\text{III}}{}_{x}(\text{OH})_2]^+[(\text{A}^n)^{\text{x/z}} \cdot \text{yH}_2\text{O}]^{\text{z-}}$ (Birjega et al., 2016). LDHs have been used extensively and are regarded as powerful adsorbents for removing anionic dyes (Lafi et al., 2016). However, due to the difficulty in separating the treated dye solution from the spent LDH, their use is restricted (Clark et al., 2019). NiAlCe-LDH successfully degraded organic dye and reached 100% (Tao et al., 2019). Moreover, LDH variant which is NiFe-LDH have been proposed for better photocatalytic performance of photocatalysts (Wu et al., 2020). Liu et al., (2016) successfully constructed an indirect photocatalytic heterostructure system comprising ZnIn$_2$S$_4$/Au/g-C$_3$N$_4$ for NO removal and CO$_2$ reduction. According to Yuliasari et al., (2022) modified Zn, Mg/Al-LDH with metal oxide as a catalyst for decomposing malachite green was reported to have superior performance to LDH.

In the present study, LDH composite Mg/Zn LDH and Mg/Zn were modified with two different polyoxometalates (POM) type Keggin K$_4[\alpha$-SiW$_{12}$O$_{40}$] and K$_3[\alpha$-PW$_{12}$O$_{40}$] and used as a photocatalyst. The investigated photodegradation factors included the impact of pH, catalyst loading, contact time, and recycled breakdown material. The selected polyoxometalate compound as intercalated with Mg/Zn LDH as intercalation due to polyoxometalate has a high negative charge that can increase the capacity for performance on degrading dye, which can make procion red photodegradation. Figure 1 shows the structure of procion red dye The XRD, FTIR, SEM, and UV-DRS were used to characterize the prepared and synthesized material.

**EXPERIMENT**

**Chemicals and instrumentation**

The study made use of sodium phosphate, sodium tungstate, sodium carbonate, magnesium nitrate, zinc nitrate, aluminum nitrate sodium hydroxide, and hydrogen chloride. One of the
synthetic dyes, procion red mx-5B, has formula C_{67}H_{44}Cl_{12}N_{16}Na_{14}O_{10}S_{2}, and maximum absorbance at λ_{max} 615 nm. A Rigaku XRD Miniflex-6000 diffractometer was used to characterize the materials. Shimadzu FTIR Prestige-21 performed FTIR analysis. The UV-Vis Biobase BK-UV 1800 PC spectrophotometer was used to measure the degradation of PR between 660–668 nm. Diffuse UV-Vis Reflectance Spectrometer JASCO V-760 was used for band gap analysis, while SEM FEI Quanta 650 was used for morphology analysis.

**CATALYST SYNTHESIS AND PREPARATION**

**Synthesis of Mg/Zn LDH**

LDH precursors eventually achieved the layered structure after being agitated in water with water combining into the structure to form the LDH phase (Wang, 2016). Both Mg and Zn LDH were generated using a modified co-precipitation. The details are as follows: an amount of magnesium nitrate 18.75 g was mixed in water and aluminum nitrate 9.3 g was dissolved in a water stirrer for 2 hours. Sodium hydroxide was added to the mixture to bring the pH level reach to pH 10. This mixture was stirred for 6 hours at 85 °C; then, MgAl LDH was weighed. A two-step synthesis of ZnAl-LDH by adding 0.25 M aluminium nitrate and 0.75 M zinc nitrate, the addition of sodium hydroxide to reach pH 8, then stirred up to 18 hours at 85 °C.

**Preparation of catalyst composite**

Mg/Zn LDH is modified with a polyoxometalate compound, Keggin K_{[α-SiW_{12}O_{40}]} and K_{[α-PW_{12}O_{40}]}]. The composite was mixed by adding 1 g polyoxometalate and 2 g LDH with adding sodium hydroxide 1 M. The suspensions were created fast for two days while being exposed to N_{2} gas for 2 days. Then, the suspension dried after being washed at a temperature of 80°C for 12 hours. XRD, FTIR, SEM, and UV-DRS were used to characterize materials.

**Photocatalytic activity**

In a 20 mg/L procion red, the initial step that was to place it under the dark conditions and agitate for 30 minutes on magnetically a desorption equilibrium; then, the photocatalytic activity of the sample was assessed. The photodegradation process applied at degradation optimization involves pH range fluctuations (3, 5, 7, 9, and 11), catalyst loading at 0.02, 0.04, 0.06, 0.08, and 0.1 g, and for the contact time degradation at 10, 20, 30, 40, 50, 60, 70, 80, 90, 100, 110, and 120 minutes. Utilizing UV light, this photodegradation process was carried out. The following equation formula is used to define the percentage of degradation (%) = (Co – Ct)/Co × 100, Co is the dye concentration at the start of the degradation process, and Ct is the dye concentration after degradation has finished (Hadjadjev-Kostic et al., 2017).

**Regeneration experiment**

To determine the photocatalyst repeatability, the solution and the suspension were separated by a precipitate. The precipitate powder was then dried for 24 h at 70 °C. The photocatalytic reaction was carried out on the solid powder. To confirm the reproducibility of both the pristine LDH and LDH composite photocatalyst, the above method was carried out a fifth time.

**RESULTS AND DISCUSSION**

**Analysis and characterization of catalyst**

The XRD patterns of the prepared MgAl-LDHs, ZnAl-LDHs, MgAl-PW_{12}O_{40}, MgAl-SiW_{12}O_{40}, ZnAl-PW_{12}O_{40}, and ZnAl-SiW_{12}O_{40} were shown in Figure 2. As shown in Figure 2, the diffraction peak of both pristine LDH MgAl-LDHs and ZnAl-LDHs can be detected at approximately 2θ = 10.39°, 20.17, 34.6 and 60.32 all coincide with the LDHs structure which is similar to the crystal plane (003), (006), (009) and (110). The crystallinity of the material is pristine, as evident by the sharp diffraction peaks. The XRD pattern of the material composite is shown in Figure 2. The peaks at 2θ and crystal plane = 11.80 (003), 23.59 (006), 31.90 (009), 46.74 (015), and 61.7 (110), respectively, for MgAl-PW_{12}O_{40} the same kind LDH which is MgAl-LDH intercalated SiW_{12}O_{40} showed diffraction peak at 8.41, 25.07 and 34.60. Furthermore, composited material on ZnAl-PW_{12}O_{40} showed a bit more peak diffraction than ZnAl-SiW_{12}O_{40}. ZnAl-SiW_{12}O_{40} observed at 2θ = 8.16, 25.27, 33.8, and 66.3, corresponding to the crystal planes (003), (006), (015), and (110). ZnAl-PW_{12}O_{40} showed at peak 7.73, 28.6 and 35.6. In this research, it was found that the involvement of polyoxometalate is supported in LDH layers.
FTIR characterization

Raman spectrum was performed to analyze the molecular vibration of pristine LDH, LDH composite, and polyoxometalate compound. The O-H stretching vibration corresponds to the absorption band at 2917 cm⁻¹. The vibration of H₂O showed 1735 cm⁻¹. The value of 1399 cm⁻¹ corresponds to CO₃²⁻. The low wavenumber from 400 – 800 cm⁻¹ can be attributed to the metal hydroxide sheets in the brucite-like lattice (M-O and O-M-O) (Li et al., 2018). The FTIR spectrum of polyoxometalate contains bands at 925-789 cm⁻¹ (W-O-W), 979 (W=O), 1020 cm⁻¹ (W-O) (Lesbani & Mohadi, 2014). Mirzaei et al., (2019) interpreted α-Keggin SiW₁₂ at bands 813, 882, 927, and 973 cm⁻¹ for W-O. LDH intercalated polyoxometalate compound, the spectrum of the LDH composite showed the successfully modified by showing the absorption band at 794 – 547 cm⁻¹.

Figure 3 shows that generally, LDH composites were maintained by being supported with polyoxometalate. The characteristic vibration disappeared, whereas the vibration of LDHs was preserved; it indicated that the layered structures of polyoxometalate were dissolved.

Structural analysis

Figure 4 shows the representative SEM image of LDH composite and pristine LDH can be seen with morphology consisting of dominant nanoparticles, noticeably the aggregation of the LDH composite happens, and the EDX of both pristine material and material composite is shown in Table 1. Pristine material and material composite indicates the atomic ratio of LDH and polyoxometalate of 2:1. The SEM image of MgAl-SiW₁₂O₄₀ is shown in Figure 4e the
### Table 1. EDX of pristine material

<table>
<thead>
<tr>
<th>Element</th>
<th>MgAl-HLG (%wt)</th>
<th>ZnAl-HLG (%wt)</th>
<th>K₂SiW₁₂O₄₀ (%wt)</th>
<th>K₂PW₁₂O₄₀ (%wt)</th>
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### Table 2. EDX of material composite

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<th>Element</th>
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<th>MgAl-Si (%wt)</th>
<th>ZnAl-Pw (%wt)</th>
<th>ZnAl-Si (%wt)</th>
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<td>Mg</td>
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**Figure 4.** SEM Analysis of MgAl-LDH (a) ZnAl-LDH (b) MgAl-SiW₁₂O₄₀ (c) MgAl-PW₁₂O₄₀ (d) ZnAl-SiW₁₂O₄₀ (e) and ZnAl-PW₁₂O₄₀ (f)
polyoxometalate appears in the small particle that sticks on the surface of MgAl-LDH. The material in this study shows a small size with a heterogeneous shape and aggregate formation. In turn, the EDX spectrum in Table 1 and Table 2 of the prepared material revealed the atomic ratio of Mg and Zn expected the composite resulting in the same amount of precursor was used; the atomic percentage of O is almost the same after being composed LDH with polyoxometalate.

**Effect of catalyst dosage**

The effect of the pristine material (MgAL-LDH and ZnAl-LDH) and composite material (MgAl-PW$_{12}$O$_{40}$, MgAl-SiW$_{12}$O$_{40}$, ZnAl-PW$_{12}$O$_{40}$, and ZnAl-SiW$_{12}$O$_{40}$) dosage (0.02; 0.04; 0.06; 0.08 and 0.1 g) on the degradation efficiency of procion red was investigated (as shown in Figure 5) the initial concentration of procion red was 20 mg/L and for the pH of the solution was not adjusted. It can be seen that the degradation was efficiently enhanced by the increased catalyst dosage, based on Figure 5, the highest increase reached 91% of procion red degradation. It is because the more catalyst added the more radicals can be generated for photodegradation dye. Thus, the photocatalytic can efficiently involve.

**Effect of initial pH**

The surface characteristics of the catalyst are affected by the pH level of the solution; this can enhance the ionization of procion red dye and the production of active radicals. The original pH value of the solution was pH, 3, 5, 7, 9, and 11 with a 20 mg/L catalyst dosage. Figure 6 illustrates how the degrading effectiveness and reaction rate constant varied as the initial pH increased. At pH 1, the removal effectiveness increased for MgAl-Si
Effect of reaction time

The analysis and evaluation of the synthesized material’s photocatalytic capability due to the degradation of reaction time under visible light. Figure 7 presents the results of the pristine material and composite material, where composite material exhibits great photodegradation efficiency. 20 mg/L of catalyst was employed in the solution to study the effects of time degradation at 2 hour reaction periods. The higher value of converting the procion red dye on the degradation process was determined by detecting an increase in reaction time. The highest reaction time was extended for one or more hours, the more efficient degradation process, which is the amount of procion red degraded. The percentage degradation of procion red of MgAl-LDH, ZnAl-LDH, ZnAl-[PW_{12}O_{40}], ZnAl-[SiW_{12}O_{40}], MgAl-[PW_{12}O_{40}], MgAl-[SiW_{12}O_{40}] was 68%, 70%, 56%, 79%, 74%, and 80%, respectively.

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of procion red by heterogeneous photo-Fenton process. Journal of Environmental Chemical Engineering, 8(4).


