Removal of Dyes by Adsorption Process Using Date Pits as Material Environmentally Friendly

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ABSTRACT
This study is based on the use of a natural material in the adsorption process to remove organic pollutants. The objective is to assess its effectiveness in adsorbing the organic pollutant MB from an aqueous solution, while operating in an open system. The DP bioadsorbent was characterized using FTIR and SEM. To determine their effect on adsorption efficiency, a number of variables were examined, including contact time, concentration of pollutant MB, adsorbent mass, pH, temperature, and adsorbent particle size. The effect of these variables on adsorption efficiency shows that a removal rate of 92.66% is achieved under optimum conditions, including a contact time of 35 minutes, a concentration of pollutant MB of 22.5 mg·l⁻¹, an adsorbent mass (m_DP) of 1.1 g·l⁻¹ and a solution pH of 5.6. In addition, a progressive decrease in adsorption efficiency is observed with increasing temperature and adsorbent mass. On the other hand, this efficiency increases with increasing a concentration of pollutant MB.

Three popular models, the Freundlich, Langmuir, and Dubin-Radushkevich models, have been used to examine the adsorption isotherms of the MB dye on DP. With a correlation factor of 0.98, it was discovered that MB adsorption monitored by the Freundlich isotherm. The Langmuir and Dubinin-Radushkevich models, however, do not adequately describe the data. The kinetic results were studied using the pseudo-first-order and pseudo-second-order equations, and show that MB dye adsorption on DP (adsorbent) follows the pseudo-second-order model. Also estimated were thermodynamic parameters such as (ΔH°), (ΔS°), (ΔG°), enthalpy, entropy, Gibbs free energy respectively to anticipate the character of adsorption. The results indicate that the adsorption process of MB on the bioadsorbent is exothermic. The results derived from the ΔG° values lead to the conclusion that the adsorption of MB occurs spontaneously.

Keywords: adsorption, BM, natural adsorbent, DP, isotherm, open system, kinetics, thermodynamics.

INTRODUCTION

The term “pollution” refers to the degradation of the environment brought on by artificial substances, which has resulted in the extinction of numerous animal and plant species as well as the rise of new phenomena that are damaging to human health, such as climate change (Al-Ghouti et al., 2010). In addition to impacting the air and soil, this pollution also has a significant impact on a considerable portion of the water. Various industrial sectors, including textiles, paper, leather, food, and cosmetics, utilize colorants extensively (Ait Hmeid, 2021). These industries also use a tremendous amount of water. Once discharged, these colors seriously harm human health by having mutagenic and cancerous consequences (Alakhras et al., 2020; Alipour et al., 2021), various modifications to the aquatic environment (Adachi et al., 2023) if they are released into the environment without being properly treated or with inadequate treatment (Ahmad et
Several methods for treating dye-laden effluents have been developed and tested in order to lessen the effects of this pollution. These include biological processes (Ledakowicz et al., 2001), coagulation/flocculation (Zhou et al., 2014), photo degradation (Diarra, 2019), ozonation (Soares et al., 2006), oxidation (Adachi et al., 2022), separation through membranes (Tan et al., 2015; Alasadi, 2019). Adsorption is thought to be one of the most successful techniques that has been used to successfully remove colors from wastewater (Abdallah et al., 2016). Because it is inexpensive and easily accessible, the facility to easily eliminate dyes from Water solutions by means of various materials, in particular activated carbon (Miyah et al., 2015), has been the subject of extensive research (Douara et al., 2016; Susanti et al., 2019). In recent times, a variety of adsorbents have been developed for the purpose of eliminating dyes (Zhang et al., 2016). Examples of such materials include: clay (Awasthi et al., 2019), double hydroxides in layers (Shan et al., 2015), steel oxides (Chakma and Moholkar, 2016), changed-nature goethite (Ouachtak et al., 2020), sedimentary materials (Dra et al., 2019). Activated carbon is an incredibly expensive product. It is therefore exciting to look for an adsorbent that is less expensive, more effective, and natural (Arami et al., 2005). Alternative materials for dye adsorbents have been suggested, including bio-materials. Date pits in particular have drawn a lot of interest because of their advantages, including affordability, availability, and environmental safety. Date pit fruit (DP) consists of seeds and a pulpy pericarp. A byproduct of many date processing plants, date palm seeds are used to make pitted dates, date powder, date syrup, date juice, chocolate-covered dates, and date confections (Rahman et al., 2007). Date palm pits are also numerous and widely dispersed, which makes them attractive environmental adsorbents that can be used in industrial operations (Saad et al., 2008). In a study, Javid et al. used modified carbonized DP with ZnO nanoparticles to remove bisphenol A and nonylphenol from aqueous solutions. Under optimal conditions, they discovered that the maximum removal efficiency was 95% (Javid and Malakootian, 2017; Javid et al., 2019). Nevertheless, the thorough investigation of date pits’ ability to adsorb BM dye has not yet been conducted (Arami et al., 2005). Therefore, the aim of this work is to use a variety of techniques, including SEM and FTIR, to assess the physical and chemical characteristics of bioadsorbent. The effectiveness of employing date pits as an economical and environmentally acceptable natural bio adsorbent to remove MB dyes from aqueous media will then be assessed. The effects of contact time, initially MB concentration, bioadsorbent mass (DP), solution pH, temperature (T), and DP particle size on the MB adsorption process on the bioadsorbent were also studied.

MATERIALS AND METHODS
PREPARATION OF A BIOADSORBENT

De-pitted (DP) Moroccan dates had their pits recovered, separated, and rinsed to get rid of any contaminants. They were then baked for 24 hours at 110°C to dry them out (Ali et al., 2022). Subsequently, the bio adsorbent was ground in a grinder and sieved to obtain particles of the same size with diameters of 40, 63, 125, and 200 μm (Fig. 1). The pits of dates consist of approximately hemicellulose, lignin, cellulose, and carbohydrates (Pal et al., 2021; Bouchelta et al., 2008).

Adsorbate

The dye investigated in the research is Methylen Blue (MB) of analytical-grade, which was obtained from Sigma-Aldrich. Its chemical formula is provided, and the Table 1 presents the discussed maximum absorption wavelength. The principal drawbacks of dye (MB) in textile wastewater are its non-biodegradability, its high visibility due to its bright color even at extremely low dye concentrations (1 mg·L⁻¹), and its high toxicity, which makes it difficult to remove using conventional procedures (Bayomie et al., 2020).

Adsorption study

For this study, a sample solution was created from the BM dye. In a batch system with continuous stirring, the adsorption study was carried out using 1 g of date pits (DP) combined with the BM solution at room temperature (Fig. 2). To investigate the kinetic adsorption, the concentration was determined using a UV-visible spectrophotometer (VR, 2000) at a wavelength of 665 nm. Every five minutes, several samples were obtained. The sample was then centrifuged to separate the bioadsorbent (DP) from the MB liquid. During the adsorption experiment, the pH of the solution was
adjusted using 0.2 M hydrochloric acid (HCl) and 0.2 M sodium hydroxide (NaOH) from Sigma-Aldrich. BM dye removal efficiency was calculated using the following formula (El Ouadrhiri et al., 2023):

\[
\text{Removal (\% R)} = \left( \frac{C_0 - C_t}{C_0} \right) \cdot 100
\]

where: \( C_0 \) and \( C_t \) represent, respectively, residual MB concentrations at \( t = 0 \) and \( t \neq 0 \).

The following equation was used to determine the date pits’ capacity to adsorb MB dye (Narasimharao et al., 2022):

\[
q_e = \left( \frac{C_0 - C_e}{m} \right) \cdot V \tag{2}
\]

In this context, \( q_e \) (mg·g\(^{-1}\)) corresponds to the equilibrium adsorption capacity, \( C_0 \) (mg·L\(^{-1}\)) indicates the concentration of Dye (MB) at \( t=0 \), \( C_e \) (mg·L\(^{-1}\)) is the concentration of MB dye at equilibrium, \( V \) (L) denotes the volume of the MB solution, \( m \) (g) represents the mass of DP.

**RESULTS AND DISCUSSIONS**

**Characterization of FTIR**

Figure 3 displays the infrared spectra (500–4000 cm\(^{-1}\)) of the powdered date pits. The chemical analysis of the DP powder reveals wide bands in the range of 3650–3000 cm\(^{-1}\), attributed to the stretching of O-H bonds (Wang et al., 2016). The C-H stretching vibrations of -CH\(_3\) and -CH\(_2\) could be responsible for the peaks seen at 2928, 2850 cm\(^{-1}\), respectively. The presence of proteins in the date wells can be verified by the picks observed at 1643 and 1545 cm\(^{-1}\), which are attributed to Amide I and Amide II respectively (Yuvaraja et al., 2020). The existence the alcohol is confirmed by the strong band at 1400 cm\(^{-1}\), which represents the vibration of C-H deformation, the band at 1605 cm\(^{-1}\), which corresponds with O-H in the plane, and the band at 1000 cm\(^{-1}\) (Dutta, 2017). Analysis of this graph indicates that the material under study is primarily composed of carbon, hydrogen, and oxygen atoms, suggesting it to be an organic compound.

**Characterization of SEM**

SEM was utilized to visualize the texture and morphology of the natural adsorbent of DP in its raw state 4(a), 4(b), and after conducting the BM adsorption tests 4(c), 4(d). The figure below 4(a), 4(b) shows that our material has a smooth and

<table>
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<th>Organic pollutant</th>
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<th>Molecular mass (g/mol)</th>
<th>Maximum length (nm)</th>
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<td>( C_{16}H_{18}ClN_3S )</td>
<td>319.85</td>
<td>665</td>
<td>Cationic dye</td>
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**Fig. 1.** Illustration of date pits adsorbent preparation.
porous structure, indicating a good potential for trapping and adsorbing the BM dye on the surface of the adsorbent (date pits) (Hassan et al., 2020). However, after BM adsorption on the date pits, SEM observation reveals that the surface of the BM-loaded material appears rugged and striated due to pore surface coverage by adsorbed BM dye molecules in figures 4(c) and 4(d). This finding suggests that date pits might work well as a bio-adsorbent to remove dangerous dyes from wastewater.

Study of different parameters on elimination rates effect of contact time

The influence of contact time improves the kinetic performance of adsorption between the adsorbent and the adsorbate as well as the equilibrium adsorption of pollutants (Gómez et al., 2007). The results of evaluating the BM adsorption by date pits at various time intervals are shown in Figure 5. As illustrated in Figure 5, during initial 35 minutes of the experiment, the concentration of BM adsorbed on the material increases with time, which perhaps be attributed to the availability of activated surfaces on the bioadsorbent (Yang, 2011). From 35 to 55 minutes, no significant change in the concentration of BM is observed. The results indicate that adsorption equilibrium is reached after only 35 minutes with a capacity of 14.56 mg/g. Perhaps equilibrium is reached after the majority of sites are occupied by the organic pollutants of BM (Li et al., 2009). For further experiments, it is recommended to fix the contact duration at 35 minutes, which is considered the optimal time.

Effect of MB concentration

According to the curves in Figure 6, a quick accretion of capacity is generally observed at low concentrations, and this capacity continues to increase with higher initial concentrations of MB (Ahmad et al., 2022). This is due to the high affinity between the exchange bioadsorbent and dye. A saturation plateau then appears, indicating saturation of the active sites on the adsorbent material. In summary, adsorption efficiency increases from 5.88 mg·g⁻¹ to 24.63 mg·g⁻¹ as the concentration of MB organic pollutants rises from 15 mg·L⁻¹ to 35 mg·L⁻¹. Figure 7 shows the variations of adsorption capacity and removal efficiency with MB dye concentrations. The two trajectories are observed to cross at a location that corresponds to the optimal concentration (Pandian et al., 2015). The optimum initial concentration of the MB solution was determined to be 22.5 mg·L⁻¹ based on the results. This concentration achieved a decolorization rate of approximately 91% with an adsorption efficiency of 21 mg·g⁻¹.
The proportion of bioadsorbent applied plays a crucial role in the adsorption procedure, as this directly affects the efficiency of dye removal (Xing et al., 2023). Figure 8 shows variations in adsorption capacity (qₑ) and discoloration rate (%R) depending on bioadsorbent mass (DP). The results demonstrate that increasing the mass of date pit adsorbent from 0.5 to 2.5 g results in an increase in the decolorization rate from 86.22% to 89.65% for organic MB pollutants (Adachi et al., 2023). Conversely, there is a decrease in adsorption efficiency, decreasing from 40.88 mg·g⁻¹ to 11.84 mg·g⁻¹. This observation can be explained by the fact that increasing the bio adsorbent load...
increases the number of active adsorption sites (El Qada et al., 2006). Thus, the availability of adsorption sites contributes to enhancing the retention efficiency. Based on the intersection of the adsorption capacity and decolorization rate curves, it was determined that the optimal mass of bio adsorbent for improved adsorption in the removal of BM dyes is approximately 1.1 g. At this mass, a decolorization rate of approximately 77.5% is observed, with an adsorption efficiency of 21.5 mg·g\(^{-1}\).

**Effect of pH**

To maximize contaminant adsorption by functional adsorbents, solution pH is a crucial parameter (Tor and Cengeloglu, 2006). Figure 9 illustrates the impact of pH on MB pollutant removal for a pH interval of 2 to 12. Date pits remove less dye at acidic pH levels for of the electromagnetic force among positively charge elements the bioadsorbent and the MB pollutant larger pH levels were observed to result in a larger percent (%) of dye removal by the bio adsorbent due to the electromagnetic attraction between the charge elements of the DP and the MB organic pollutant. Larger pH levels were observed to result in a larger percent (%) of MB pollutant removal by DP due to the electromagnetic force between the charged molecules DP and the MB atoms (Senthilkumaar et al., 2005). The results in the figure below demonstrate that a basic environment is more favorable for dye adsorption on the studied material (DP) compared to an acidic environment. This can be attributed to the fact that the addition of H\(^+\) protons neutralize the negative charge of the material, which hinders the adsorption of dyes in highly acidic environments and, conversely, favors it in basic environments.

**Effect of temperature**

In the 300–350 K° range, the effect of T (K°) on MB adsorption capability was studied. Figure 10 presents the impact of T (K°) on the dye retention rate on the adsorbent. From the figure, we observe that this rate decreases with increasing reactor temperature (Rahmani et al., 2010). According to this phenomenon and the Arrhenius equation, the surface reaction is exothermic, and each rise in temperature slows it down. The analysis of this figure shows that increasing the temperature leads to a slight decrease in adsorption capacity. The optimal adsorption occurs at room temperature.
Effect of bioadsorbent size

Date pits (DP) that had been dried up and were still raw were broken up and sifted into various split sizes (250–355–500–700 um) to examine the effects of particle size on the adsorption process (Banat et al., 2003). Figure 11 demonstrates how the dye’s adsorption ability declines as particle size rises. This is due to the fact that adsorption is reduced as particle size increases because of a reduction in specific surface area. It is possible to confirm that optimum adsorption is achieved at a particle size of 225 μm.

**ADSORPTION ISOTHERMS**

**Langmuir isotherms**

The Freundlich, Langmuir and Dubin-Radushkevich models for the analysis of adsorption isotherms were reviewed and used to characterize the adsorption process for our research results. One of the theories for explaining monolayer adsorption is the Langmuir isotherm. It presupposes a homogenous adsorption surface with identically energetic binding sites. The following is how to express the Langmuir isotherm’s linear form (Da, 2001):

\[
\frac{1}{q_e} = \frac{1}{q_{\text{max}}} + \frac{1}{(K_L \times q_{\text{max}})} \times \frac{1}{C_e}
\]  

The maximal adsorption capacity under experimental conditions is represented by \(q_{\text{max}}\) (mg/g), \(q_{\text{max}}\) and \(K_L\) are calculated from the plot of \(C_e / q_e\) versus \(C_e\).

The expression of the dimensionless parameter \(R_L\), which represents the equilibrium or adsorption intensity, can be derived from further analysis of the Langmuir equation.

\[
R_L = \frac{1}{1+K_L C_0}
\]

The \(R_L\) parameter, which is based on the initial amount of adsorbate \(C_0\) measured in mg/L, is considered a more reliable indicator of the adsorption process (Figure 12). The \(R_L\) value can be classified into four categories: (i) \(R_L\) between 0 and 1 indicates good adsorption, (ii) \(R_L\) above 1 indicates bad adsorption, (iii) \(R_L\) equal to 1

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**Fig. 10.** Impact the T(K°) on adsorption of MB \((t = 35 \text{ min}, \text{pH} = 5.6, [MB] = 22.5 \text{ mg/l}}, V = 50 \text{ ml}, m_{DP} = 1.1 \text{ g})\).

**Fig. 11.** Impact of particle size on MB adsorption on DP \((t = 35 \text{ min}, T = 25°C, \text{pH} = 5.6, [MB] = 22.5 \text{ mg/l}}, V = 50 \text{ ml}, m_{DP} = 1.1 \text{ g})\).

**Fig. 12.** Langmuir adsorption isotherm \((t = 35 \text{ min}, T = 25°C, \text{pH} = 5.6 [MB] = 22.5 \text{ mg/L}}, V = 200 \text{ ml}, m_{DP} = 1.1 \text{ g})\).
represents linearly adsorption, and (iv) $R_L$ equal to 0 means non-reversible adsorption (Angove et al., 1997).

**Freundlich isotherm**

The logarithmic equation applied to the investigation of the Freundlich isotherm (Fig. 13) is as follows (Boparai et al., 2011):

$$\log (q_e) = \log (k_F) + \frac{1}{n} \log (C_e)$$  \hspace{1cm} (5)

A curve with a slope of $1/n$ and an origin ordinate of $\log (k_F)$ is obtained by transporting a log $(q_e)$ as a function of $C_e$. This allows us to calculate the constant $K_F$ (mg/g) and the heterogeneity factor $1/n$.

**Dubinin-Radushkevich isotherm**

In contrast to the Langmuir model, Dubinin-Radushkevich (Fig. 14) requires a homogenous surface. Its theory of micropore volume filling is predicated on the idea that the adsorption potential is changeable and that the adsorption free energy is correlated with the level of pore filling (Vijayaraghavan et al., 2006). The following equation results in the Dubinin-Radushkevich isotherm (Ayawei et al., 2015):

$$\ln(q_e) = \ln (q_m) - K_{DR} \varepsilon^2$$ \hspace{1cm} (6)

where: $q_m$ is the equilibrium adsorption capacity of DP (mg·g$^{-1}$), $K_{DR}$ presents the Dubinin-Radushkevich constant (mol$^2$·Kj$^{-2}$), and $\varepsilon$ ($\varepsilon = RT \ln (1 + 1/C_e)$).

The average sorption energy ($E$) in kJ·mol$^{-1}$ can be determined by utilizing the following equation (Saltali et al., 2007).

$$E = \frac{1}{\sqrt{-2K_{DR}}}$$ \hspace{1cm} (7)

As far as applied models are concerned, both the Dubinin-Radushkevich and Langmuir adsorption models showed very low correlation coefficients for the experimental data, $R^2 = 0.70$, $R^2 = 0.83$, respectively. Table 2 demonstrates the inability of these models to accurately describe the adsorption process of MB organic pollutant on the studied biosorbents. The inapplicability of the Dubinin-Radushkevich and Langmuir adsorption isotherms suggests that adsorption occurs on non-uniform adsorption sites due to the inequality of adsorption site activities on different date pit-based biosorbents. Correlation factor results presented in Table 2 demonstrate that the Freundlich model is the most appropriate for characterizing the adsorption of the organic pollutant MB on the biosorbent DP, as shown by the highest $R^2$ correlation coefficients. The parameter $1/n$ represents the relative distribution of energetic adsorption sites and depends on both the nature and intensity of the adsorption process. The obtained value for $1/n$ is typically less than 1, indicating
that the adsorption of MB organic pollutant on DP is favorable and corresponds to a chemical adsorption process.

**Adsorption kinetics**

The following equations define the kinetics of the pseudo-first-order and pseudo-second-order models, respectively (Fan et al., 2003):

\[
\ln(qe-qt) = \ln(qe) - \frac{k_1}{2.303} t \\
\]

\[
\frac{t}{q_t} = \frac{1}{K_2 qe^2} + \frac{1}{qe} t \\
\]

Plotting \( \ln(qe-qt) \) as a function of time (Fig. 14) results in a line with a slope of \( k_1 \) if the Lagergren connection is confirmed. A line with a slope of \( 1/q_e \) and a y-intercept of \( 1/k_2 qe^2 \) must be obtained in order to plot \( t/q_t \) as a function of time (Fig. 15). Table 3 shows that the pseudo-first-order model’s equation is not linear and has a very low coefficient of correlation \( R^2 \), which shows that the experimental absorption capacity differs greatly from the value predicted by this model. Thus, we can deduce that the adsorption kinetics does not follow the pseudo-first-order model (Fig. 15). However, from the obtained results (Fig. 16, Table 3), it is evident that the variation of \( t/q_t \) with time is highly linear, and the regression coefficient \( R^2 \) is satisfactory. Thus, we can draw the conclusion that the kinetic adsorption of MB using date pits follows the pseudo-second-order model, which implies that the chemical interactions between dye molecules and the adsorbent surface are characterized by electron transfers (Krishnamoorthy et al., 2019).

**Thermodynamic study**

To understand the adsorption phenomenon better, knowledge of the thermodynamics of adsorption is essential (Fig. 17). Based on experimental data and the accompanying formulae, the Van’t Hoff equations have been used to calculate the thermodynamic variables the adsorption process, including three variations such as: \( (\Delta G^0) \), \( (\Delta H^0) \), \( (\Delta S^0) \), Gibbs free energy, enthalpy, entropy \( (\Delta S^0) \), respectively (Singh and Pant, 2004),

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Fig. 15. Pseudo-first-order kinetics of MB adsorption on DP (\( t = 35 \) min, \( T = 25^\circ C, \) pH = 5.6, \( [MB] = 22.5 \) mg/l, \( V = 50 \) ml, \( m_{DP} = 1.1 \) g)

Fig. 16. Pseudo-second-order kinetics of MB adsorption on DP (\( t = 35 \) min, \( T = 25^\circ C, \) pH = 5.6, \( [MB] = 22.5 \) mg/l, \( V = 50 \) ml, \( m_{DP} = 1.1 \) g)
or $\Delta G^\circ$ (kJ/mol), T (K), $\Delta H^\circ$ (kJ/mol), R (J/mol·K), $\Delta S^\circ$ (J/K) present the standard free energy, the absolute temperature, the universal constant, the standard entropy respectively.

The negative values of $\Delta G^\circ$ at 298, 323, 333, and 348 degrees Celsius are displayed in Table 4 and illustrate that these temperatures enhance spontaneity of adsorption. At 298, 323, and 333 K, the adsorption of MB organic pollutant onto date pit cores followed a similar pattern. The negative standard enthalpy value, $\Delta H^\circ$, for MB signifies that the adsorption process is exothermic (Sari et al., 2007). The MB molecules are more structured at the solid-liquid interface than in the liquid phase, according to the negative entropy value, $\Delta S^\circ$. The adsorption process is spontaneous, according to the negative values of the free energy, $\Delta G^\circ$.

CONCLUSION

In the current study, we prepared an adsorbent for the batch system-based adsorption of MB organic pollutant from an aqueous solution using raw date pits, a cheap and easily accessible material. The findings revealed that an equilibrium reaction took 35 minutes to complete at a pH of 10, MB organic pollutant concentration of 22.5 mg·L$^{-1}$, and an adsorbent mass of 1.1 g·L$^{-1}$. The highest adsorption efficiency was at 24 mg·g$^{-1}$. However, as temperature and particle size rose, the adsorption effectiveness fell off dramatically. Compared to other applicable isotherm models, the Freundlich isotherm model offered a superior fit for the adsorption process. The kinetic data of the MB adsorption process were closely associated with the pseudo-second-order model. The thermodynamic examination of adsorption yielded standard Gibbs free energy ($\Delta G^\circ$) and standard enthalpy ($\Delta H^\circ$) values, which show that the MB adsorption on date pits is exothermic and spontaneous. According to the study’s findings, date pits are a highly effective adsorbent for removing colors from aqueous environments.

REFERENCES


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<th>$C_0$ (mg/l)</th>
<th>$q_{exp}$ (mg/g)</th>
<th>$K_1$ (min$^{-1}$)</th>
<th>$q_t$ (mg/g)</th>
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**Table 3. Parameters of kinetic models**

**Table 4. Thermodynamic settings of the BM adsorption process on the bio adsorbent**

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<th>$\Delta H^\circ$ (kJ·mol$^{-1}$)</th>
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</tbody>
</table>

 Fig. 17. Line of $\ln(K_C)$ versus $1/T$ for adsorption of MB on DP (t = 35 min, $T = 25^\circ C$ ± 4, $pH = 5.6$, $[MB] = 22.5$ mg/l, $V = 50$ ml, $m_{DP} = 1.1$ g)


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